

Domestic ships as a potential pathway of nonindigenous species from the Saint Lawrence River to the Great Lakes

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Abstract Ballast water moved by transoceanic vessels has been recognized globally as a predominant vector for the introduction of aquatic nonindigenous species (NIS). In contrast, domestic ships operating within confined geographic areas have been viewed as low risk for invasions, and are exempt from regulation in consequence. We examined if the St. Lawrence River could serve as a source of NIS for the Laurentian Great Lakes by surveying ballast water carried by domestic vessels and comparing biological composition in predominant St. Lawrence River—Great Lakes port-pairs in order to determine the likelihood that NIS could be transported to, and survive in, the Great Lakes. Thirteen potential invaders were sampled from ballast water, while 26 taxa sampled from St. Lawrence River ports are not reported from the Great Lakes. The majority of NIS recorded in samples are marine species with low potential for survival in the Great Lakes, however two euryhaline species (copepod *Oithona*

similis, and amphipod *Gammarus palustris*) and two taxa reported from brackish waters (copepod *Microsetella norvegica* and decapod *Cancer irroratus*) may pose a risk for invasion. In addition, four marine NIS were collected in freshwater samples indicating that at least a subset of marine species have potential as new invaders to the Great Lakes. Based on results from this study, the ports of Montreal, Sorel, Tracy and Trois Rivières appear to pose the highest risk for new ballast-mediated NIS from the St. Lawrence River to the Great Lakes.

Keywords Alien species · Nonindigenous species · Invasive species · Great Lakes · Saint Lawrence River · Community composition

Introduction

The anthropogenic movement of nonindigenous species (NIS) around the world has occurred for many centuries, though the rate of invasions accelerated in recent decades owing to technological advances in transportation, expanding global economies, and attendant international trade (e.g. Mack et al. 2000; Work et al. 2005; Hulme 2009; Sylvester and MacIsaac 2010). Dispersal of NIS is strongly shaped by, and reflects movement of, transport vectors that facilitate global commerce and travel. Transoceanic ships have been a dominant vector of aquatic NIS, with ballast water, ballast sediments and hull biofouling attributed

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with introduction of hundreds of species globally (Ruiz et al. 1997; Minchin and Gollasch 2003; Bailey et al. 2007).

Vector management seeks to reduce the diversity and abundance of NIS transported to new locations (Lodge et al. 2006; Ricciardi 2006; Bailey et al. 2011a). In order to reduce invasions by ballast water, a number of countries require foreign vessels to manage ballast water by mid-ocean ballast water exchange (USCG 1998; Government of Canada 2006). Mid-ocean exchange reduces the abundance of NIS in ballast tanks through purging of biota at sea, and is particularly protective of freshwater ports through the introduction of a salinity barrier that imposes osmotic stress on freshwater NIS (Bailey et al. 2011a; Briski et al. 2013).

Ballast discharges by domestic vessels operating within the exclusive economic zone of a single country are currently exempted from regulation, despite evidence of their importance in the spread of NIS between ports (Lawrence and Cordell 2010; Briski et al. 2012b). General assumptions that regional ports within “common waters” have contiguous communities have also been called into question (Lawrence and Cordell 2010; David et al. 2013). As an example, ballast water transfers from ports on the St. Lawrence River into the Great Lakes may seem benign since the two regions are naturally connected, however, this large geographic region has distinct ecoregions with diverse biological communities where species native to one location may be nonindigenous elsewhere (Briski et al. 2012b). Ports on the Saint Lawrence River are also visited by transoceanic and coastal ships, linking Saint Lawrence River ports to more than 75 different ballast source ports in coastal North America, South America, Europe and Africa, presenting the possibility of ‘stepping stone’ invasions mediated by domestic ballast transfers to the Great Lakes (Rup et al. 2010; Bailey et al. 2011b). In fact, the Great Lakes currently support 22 NIS that are either native to, or were first reported in, the Saint Lawrence River or other North American east coast rivers, suggesting that Saint Lawrence River ports may be an important source for NIS introductions to the Great Lakes (de Lafontaine and Costan 2002; Kelly et al. 2009; Bailey et al. 2011b).

Depending on their configuration, domestic vessels are capable of hauling up to 70,000 tonnes of cargo per trip, with an average trip taking 3–4 days between

ports on the Great Lakes (Rup et al. 2010). Survival of plankton in ballast water during short transits of domestic vessels tends to be higher than that of many oceanic voyages, which can take one to several weeks (Simkanin et al. 2009, Lawrence and Cordell 2010). However, as the Saint Lawrence River has fresh, brackish and euhaline ports, uncertainty exists regarding survival of Saint Lawrence River species discharged into the Great Lakes. We expect that survival will be lowest for euhaline to freshwater movements; notwithstanding this, there are precedents of euryhaline species invading the Great Lakes (Ricciardi and MacIsaac 2000; Kelly et al. 2006; Rup et al. 2010), indicating a need for investigation of Saint Lawrence River to Great Lakes ballast water movements.

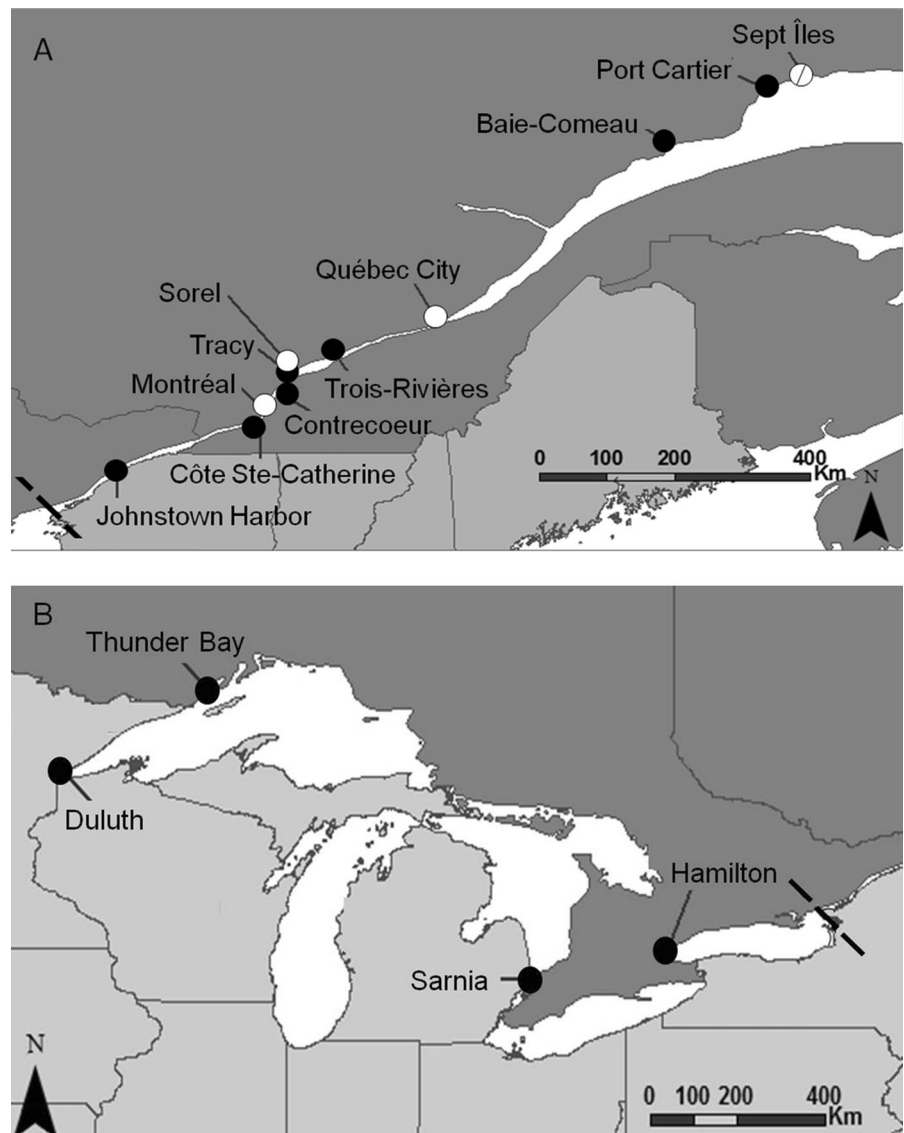
Here we explore the role of Saint Lawrence River ballast water as a vector of NIS dispersal to the Great Lakes. We surveyed Saint Lawrence River ballast water carried by domestic vessels to ports on the Great Lakes and compared biological composition in predominant Saint Lawrence River—Great Lakes port-pairs in order to determine the likelihood that NIS from the Saint Lawrence River could be transported to, and survive in, the Great Lakes.

Materials and methods

Sampling strategy

We collected ballast water samples from domestic vessel transits between the Saint Lawrence River and the Great Lakes. We also collected zooplankton and zoobenthos samples from Saint Lawrence River ports that serve as sources of domestic ballast water, as well as Great Lakes ports where Saint Lawrence River ballast is discharged (Rup et al. 2010) (Fig. 1). St. Lawrence River ports were assigned risk categories based on the amount of ballast each port sourced to the Great Lakes between 2005 and 2007, and on port salinity. The natural break in a scatter plot of total volume of Great Lakes-bound, Saint Lawrence River ballast moved by domestic ships over the 3 year period of 2005–2007 was used to assign Saint Lawrence River ports into a high or low ballast activity group. The high volume group consisted of ports (Québec City, Sorel, Montreal, Port Cartier, and Baie Comeau) that sourced more than 35,000 tonnes of ballast to the Great Lakes, while the low volume

Fig. 1 Location of ports sampled in the (a) Saint Lawrence River and (b) Great Lakes. Port marker in panel (a) reflects assigned risk categories: *black dots*—high risk; *open dots*—medium risk; *open dot with a diagonal line*—low risk. The *dotted line* in both panels indicates the separation of the Great Lakes proper and the Saint Lawrence River



group (Côte Ste Catherine, Johnstown Harbour, Contrecoeur, Tracy, Trois Rivières, and Sept Îles) each donated less than 35,000 tonnes. Next, ports were classified as fresh (0–0.5 ppt) or marine (>17 ppt), with the three ports downstream of Québec City designated marine (Baie Comeau, Port Cartier, and Sept Îles) and the remaining seven upstream ports designated freshwater (Vincent and Dodson 1999). Three risk categories were thus created for Saint Lawrence River ports that sourced ballast to the Great Lakes: high risk ports involved high volumes of freshwater (Montréal, Québec City, Sorel); medium risk ports involved a low volume of freshwater (Côte

Ste Catherine, Contrecoeur, Johnstown Harbour, Trois Rivières, Tracy) or high volume of marine water (Port Cartier, Baie Comeau); and low risk ports involved low volume of marine water (Sept Îles) (Fig. 1a).

We collected ballast water samples from 30 vessels bound for the Great Lakes that carried Saint Lawrence River ballast between May 2007 and October 2010. Nineteen of these samples were freshwater ballast (<0.5 ppt), 10 samples were marine (>17 ppt), and one was sourced from a freshwater port but appeared to be mixed with a significant volume of residual marine water (3.6 ppt). Twenty-nine samples were collected by repeatedly lowering a 53 µm vertical zooplankton

net to a depth of 2–15 m in a single location (ballast tank) until approximately 1,000 L of water was filtered. One sample was collected by pumping 50 L of ballast water from a single tank through a sounding tube, which was then passed through a 53 μm mesh sieve, since the tank hatch could not be opened. After collection, ballast water samples were preserved with 95 % ethanol and sent for taxonomic identification.

Zooplankton and zoobenthos were collected from 11 Saint Lawrence River ports on three occasions between September 2009 and September 2010, and from four Great Lakes ports in August 2010. In total, we collected 88 zooplankton and 42 benthic samples. Details regarding sampling site locations and characteristics are provided in Fig. 1 and Table 1. Zooplankton sampling in ports was conducted from a boat using triplicate vertical plankton net tows of 41 and 500 μm mesh. Upon collection, samples were rinsed from the cod-end of the net into a 41 or 500 μm sieve corresponding to the net's mesh size. Samples were rinsed while still in the sieve with 95 % ethanol to remove fine sediment before being preserved with 95 % ethanol. We collected benthic invertebrates using either a sled dredge (rocky substrate) or ponar grab (sand or mud bottom). Sled dredge samples were collected by gently dragging the net along the bottom for approximately 5 min. Ponar grab sampling

involved two to three grab samples collected from different points, which were subsequently combined into a single sample. After retrieval, benthic samples were transferred into a sieving bucket with 1 mm mesh screen bottom and washed with port water to remove sediment, clay, and organic matter. We examined macrofauna for viability (e.g., movement) with the naked eye before samples were washed into 1 L jars, preserved with 95 % ethanol, and transported to the laboratory for analysis. Four limnological parameters—temperature, salinity, conductivity and dissolved oxygen—were recorded during each port sampling event using a handheld YSI instrument (Table 1).

Animals in port plankton samples were examined under a dissecting microscope at 0.63–5 \times magnification. We split dense samples into fractions using a Folsom splitter, and each fraction thereafter examined. Animals were counted and separated into broad taxonomic groups (e.g., cladocerans, rotifers) and a maximum of 30 individuals for each group was collected for detailed taxonomic identification (Humphrey 2008). Benthic samples were rinsed with 95 % ethanol into 1 mm, 500 and 45 μm stacked sieves, with animals concentrated on each sieve subsequently counted and sorted into taxonomic groups. The 1 mm fraction was examined by eye, while the 500 and

Table 1 Recorded longitude, latitude, mean temperature, mean conductivity, dissolved oxygen, and mean salinity for Saint Lawrence River ports (September 2009, July and September 2010) and Great Lakes ports (August 2010)

Port	Longitude (°W)	Latitude (°N)	Temperature (\pm SE) (°C)	Conductivity (\pm SE) (μS)	Dissolved oxygen (\pm SE) (%)	Salinity (\pm SE) (ppt)
Baie Comeau	68.14	49.25	3.6 \pm 0.9	34 \pm 1	12 \pm 1	32.4 \pm 0.9
Contrecoeur	73.28	45.83	24.0 \pm 0.3	307 \pm 1	8 \pm 1	0.1 \pm 0.0
Côte Ste Catherine	73.59	45.41	23.1 \pm 0.0	304 \pm 3	8 \pm 0	0.1 \pm 0.0
Duluth	92.11	46.77	22.3	263	9	0.1
Hamilton	79.80	43.28	25.3	586	8	0.3
Johnstown Harbour	75.47	44.74	20.8 \pm 0.4	302 \pm 2	8 \pm 0	0.2 \pm 0.0
Montréal	73.51	45.58	24.0 \pm 0.7	307 \pm 2	9 \pm 0	0.1 \pm 0.0
Port Cartier	66.78	50.03	10.9 \pm 0.4	35 \pm 1	10 \pm 0	31.1 \pm 0.5
Quebec City	71.20	46.82	21.0 \pm 0.6	280 \pm 3	8 \pm 0	0.1 \pm 0.0
Sarnia	82.45	42.93	23.3	224	8	0.1
Sept Îles	66.37	50.19	13.2 \pm 0.3	35 \pm 1	9 \pm 0	29.5 \pm 0.5
Sorel	73.14	46.05	24.7 \pm 0.4	196 \pm 1	7 \pm 0	0.1 \pm 0.0
Thunder Bay	89.22	48.41	20.4	105	9	0.1
Tracy	73.12	46.05	23.6 \pm 0.1	320 \pm 5	8 \pm 0	0.1 \pm 0.0
Trois Rivières	72.55	46.33	24.3 \pm 0.3	293 \pm 3	8 \pm 0	0.1 \pm 0.0

45 µm fractions were examined under a dissecting microscope (0.63–5× magnification). As with zooplankton samples, we examined a maximum of 30 individuals of each broad group for detailed taxonomic identification.

Ballast water and port samples were sent out for expert identification (EcoAnalysts INC., Moscow, ID and Biologica Environmental Services Ltd., Victoria, BC, respectively), except for port rotifers. Rotifers were identified according to Koste (1978) and Stemberger (1979), and were verified by Dr. Ian Duggan (University of Waikato). The majority of port rotifers were identified only to genus level because animals were distorted owing to preservation in ethanol.

Data analysis

To assess the introduction potential posed by Saint Lawrence River ballast, we examined both the potential for arrival and potential for survival at recipient ports. We recorded the number of individuals per species and species richness for ballast water and port samples. A variety of freshwater and marine taxa were found in port and ballast water samples. Identified species were characterized as native to the Great Lakes, NIS not reported from the Great Lakes, NIS already established in the Great Lakes, or cryptic species (Stemberger 1979; Balcer et al. 1984; Kelly et al. 2009; USGS 2009). NIS not reported from the Great Lakes were further researched, primarily using the World Register of Marine Species and the Web of Science, to determine species' distribution and evaluate potential for survival in the freshwater Great Lakes. We compared the abundance of NIS not reported from the Great Lakes in our Saint Lawrence River ballast samples against similar data for vessels with foreign exchanged ballast water (Bailey et al. 2011a) to evaluate the relative importance of this pathway. The nonparametric Mann–Whitney U test with $\alpha = 0.05$ was used due to the skewed data distribution (many 'zero' values) (JMP® Version 10 statistical software; SAS Institute Inc. 2012).

Biological similarity of Saint Lawrence River—Great Lakes port-pairs was determined by comparing the community of species found in samples from individual high and medium risk Saint Lawrence River ports to that in the top two corresponding Great Lakes recipient ports. Biological similarity was calculated using species presence/absence data and

Sørensen's similarity index (SSI), using SPADE software (Sørensen 1948; Chao and Shen 2006):

$$SSI = \frac{2D_{12}}{D_1 + D_2} \quad (1)$$

where D_1 is the number of observed species in sample 1, D_2 is the number of species observed in sample 2, and D_{12} is the number of observed species in the two samples (Chao and Shen 2006). Index values ranges from zero—for communities that have no species overlap—to one, for communities that have all species in common.

Results

Total zooplankton abundance in ballast water ranged from 5.8 to 20.2×10^5 individuals m^{-3} , with a median abundance of 1.4×10^4 individuals m^{-3} (see Online Resource 1). We identified 78 distinct zooplankton taxa from ballast water samples, ranging from 4 to 26 species per sample (Online Resource 2); thirteen NIS not reported from the Great Lakes were identified from six freshwater and nine marine ballast samples. We calculated total and mean abundance of each NIS not present in the Great Lakes across all ballast water samples as a measure of the arrival potential each species exerts on the Great Lakes (Online Resource 3). Both abundance and richness of these NIS were significantly higher in marine (median 6.7×10^3 individuals m^{-3} when present; 5 NIS) than freshwater samples (median 3.3 individuals m^{-3} when present; 1 NIS) (Mann–Whitney U test; $p < 0.001$). The abundance of NIS not reported from the Great Lakes in Saint Lawrence River ballast water was also significantly greater than that sampled from foreign exchanged ballast water (median 850 individuals m^{-3} ; Briski et al. 2012b) (Mann–Whitney U test; $p < 0.01$).

Rotifers were the most abundant (69 % total abundance) and diverse (40 species; 54 % total richness) group sampled; however, all rotifer species were already present in the Great Lakes (Online Resource 2). Bivalve veligers were the second-most abundant group (23 % total abundance), followed by copepods (6 %), and cladocerans (2 %). Amphipods, cirripeds and decapods all represented <1 % of zooplankton abundance in ballast samples. Copepods were the second-most diverse group with 15 species, including eight species not reported from the Great Lakes (*Acartia hudsonica*, *Centropages hamatus*, *Epischura*

nordenskiöldi, *Temora longicornis*, *Tortanus discaudatus*, *Oithona similis*, *Coullana canadensis* and *Microsetella norvegica*) and one NIS already established (*Schizopera borutskyi*). We identified eleven cladoceran species, including one species not reported from the Great Lakes (*Evadne nordmanni*) and one NIS already established in the Great Lakes (*Eubosmina coregoni*). We identified three bivalves, including the species c.f. *Mytilus edulis* not reported from the Great Lakes, and the two dreissenid mussels already established. The two decapod taxa (c.f. *Cancer irroratus* and c.f. *Uca sp.*) found in ballast from Port Cartier are both marine species not reported from the Great Lakes. The single cirriped species (c.f. *Balanus sp.*) is a marine species not reported from the Great Lakes while the single amphipod sampled is a NIS already established (*Echinogammarus ischnus*). While the majority of potential NIS not recorded from the Great Lakes are marine species that were exclusively collected from marine ballast samples, *O. similis* is a euryhaline species reported from freshwater environments, that we also sampled in freshwater ballast from Montreal (Table 2). The marine *E. nordenskiöldi* and brackish *M. norvegica* are not reported from fresh water globally, but still exhibit potential for survival in the Great Lakes as they were sampled in freshwater ballast sourced from Montreal and Sorel (Table 2).

During the Saint Lawrence River port surveys, we identified 67 distinct taxa from freshwater ports,

including four NIS not reported from the Great Lakes (Online Resource 4; Table 2). Rotifers were the most abundant group, accounting for 46 % of total invertebrate abundance. Bivalves (veligers and adults) were the second-most abundant taxon (22 %), followed by cladocerans (20 %) and 11 other rare (12 % total) taxa (amphipods, caddisflies, copepods, dipterans, dragonflies, gastropods, mayflies, mites, mysids, oligochaetes and polychaetes). Oligochaetes were the most diverse group in freshwater Saint Lawrence River ports, with 15 species (including two not reported from the Great Lakes, *Chaetogaster diaphanus* and *Stylaria lacustris*; and one established Great Lakes NIS, *Ripistes parvasita*) followed by cladocerans (12 species, with two Great Lakes NIS, *Bythotrephes longimanus* and *Eubosmina coregoni*), rotifers (11 species), dipterans (11 species), bivalves (four species including Great Lakes NIS *Dreissena polymorpha* and *Dreissena rostriformis bugensis*), and amphipods (three species, including *Gammarus oceanicus* and *Gammarus palustris* which are not reported from the Great Lakes). Other groups included mites (three species), copepods and mayflies (two species each), and caddisflies, dragonflies, gastropods, mysids and polychaetes with one species each.

We identified fifty-two taxa from marine Saint Lawrence River ports, including 23 species not reported from the Great Lakes (Online Resource 4). Rotifers and polychaetes dominated invertebrate communities, representing 50 and 25 % of total abundance, respectively. Relative abundance of the remaining 12 major taxonomic groups was much lower, collectively accounting for 25 % of zooplankton abundance. Four taxa (polychaetes, amphipods, bivalves and dipterans) contributed most to species diversity (31, 19, 12 and 12 %, respectively), whereas the other ten taxa collectively contributed only 26 %. Only one Great Lakes established NIS, *Dreissena polymorpha*, was found in marine port samples. Taxa not reported from the Great Lakes include amphipods (*Pontogeneia inermis*, *Anonyx sarsi*, *Gammarus oceanicus*, *Caprella linearis*, *Tmetonyx cicada*, *Calliopius laeviusculus*, *Monocorophium acherusicum*, *Monoculodes tuberculatus* and *Orchomenella groenlandica*), bivalves (*Macoma calcarea*), cirripeds (*Balanus sp.*), cladocerans (*Evadne sp.*), cumaceans (*Diastylis rathkei*), decapods (*Eualus gaimardi* and *Cancer irroratus*), echinoderms (*Ophiura robusta*) and polychaetes (*Goniada maculata*, *Capitella capitata complex*, *Harmothoe extenuata*, *Gattyana*

Table 2 List of potential NIS sampled from Saint Lawrence River freshwater ballast and port samples, respectively

	JH	M	T	S	TR	QC
Potential NIS in ballast samples						
<i>Epischura nordenskiöldi</i>						x
<i>Oithona similis</i>		x				
<i>Microsetella norvegica</i>		x	x	x		
Potential NIS in port samples						
<i>Gammarus oceanicus</i>				x	x	
<i>Gammarus palustris</i>			x		x	
<i>Chaetogaster diaphanus</i>	x	x	x	x		
<i>Stylaria lacustris</i>			x		x	x

Source port of ballast water is indicated. Ports are: JH Johnstown Harbour, M Montreal, T Tracy, S Sorel, TR Trois Rivières, QC Québec City; no potential NIS were reported from Côte Ste Catherine or Contrecoeur

cirrhosa, *Nephtys cornuta*, *Pectinaria granulata* and *Harmothoe imbricata*).

We identified thirty-six species from samples collected at ports on the Great Lakes (Online Resource 5). Cladocerans and rotifers were most abundant, accounting for 37 and 34 % of total invertebrate abundance, respectively. Eight other taxa collectively accounted for 29 % of total individuals surveyed. Cladocerans and oligochaetes exhibited greatest diversity with 11 and nine species, respectively, followed by rotifers (eight species) and the remaining five taxa (11 species). Two established cladoceran (*Bythotrephes longimanus* and *Eubosmina coregoni*) and dreissenid (*Dreissena polymorpha* and *Dreissena rostriformis bugensis*) NIS were recorded in port samples.

Sørensen's similarity index values for all high and medium risk donor Saint Lawrence River ports and their respective top Great Lakes recipient ports ranged from 0.07 (Port Cartier-Thunder Bay) to 0.53 (Tracy-Thunder Bay), indicating that none of the Saint Lawrence River—Great Lakes port-pairs are highly similar in community composition.

Discussion

This is the first study to examine the potential for introduction of new NIS through domestic ballast water transported from the Saint Lawrence River to the Great Lakes. The abundance of NIS not yet present in the Great Lakes was significantly higher for ships with marine versus freshwater ballast, however no new freshwater NIS not reported from the Great Lakes were found in any Saint Lawrence River ballast water or port sample. We did, however, record 33 marine species and two brackish water taxa not reported from the Great Lakes, seven of which were collected from freshwater samples. In addition, we recorded the euryhaline copepod *O. similis* from multiple ballast water samples. We observed only minimal overlap in community composition of ballast and port samples (12 shared taxa), indicating that additional species might be detected with increased sample effort (see also Adebayo 2012). The list of potential NIS reported in our port and ballast water samples is only an indication of risk, as our sampling was limited in both spatial and temporal coverage, and we were unable to identify all organisms to species level, making our species list non-exhaustive. Other studies have shown that the Saint Lawrence River is inhabited by

at least one freshwater invertebrate NIS (spiny cheek crayfish, *Orconectes limosus*) and two freshwater fish NIS (the tench *Tinca tinca* and cutthroat trout *Oncorhynchus clarki*) that are not reported from the Great Lakes, thus there is at least some potential for freshwater NIS to be moved from the Saint Lawrence River to the Great Lakes (de Lafontaine and Costan 2002).

Given that we observed seven marine species from freshwater ballast or port samples, a potential exists for the establishment of marine Saint Lawrence River taxa at ports on the Great Lakes. Although invasion potential of marine taxa into freshwater habitats is generally considered low, the establishment risk of such species in the Great Lakes should not be discounted (Drake and Lodge 2007; Briski et al. 2012a). While not common, some invertebrates are able to successfully transition from marine to freshwater habitats (Lee 1999). For example, the marine copepod *Eurytemora affinis* had achieved multiple, independent freshwater transitions with strong selection pressure for freshwater tolerance before successful invasion in the Great Lakes (Lee 1999; Lee et al. 2003). Similarly, the marine amphipod *Gammarus tigrinus* has several independent lineages that have invaded freshwater habitats, with the introduced population in the Great Lakes likely the product of intense selection following ballast-mediated introduction (Kelly et al. 2006).

Our results indicate that four ports (Montreal, Sorel, Tracy and Trois Rivières) on the Saint Lawrence River pose the highest risk for ballast-mediated transfer of species to the Great Lakes. These four ports are also important recipients for ballast water discharges, receiving a combined annual total of 191,692 and 424,571 tonnes of unexchanged coastal and international ballast, respectively, thus underscoring that they could also serve as intermediate stages of 'stepping stone' or secondary invasions to the Great Lakes (Floerl et al. 2009; Bailey et al. 2011b). The presence of established NIS in 90 % of our ballast water samples supports the hypothesis that domestic vessels facilitate secondary invasions of NIS within the Great Lakes. Furthermore, translocation of established NIS can introduce novel genotypes to populations already established in the Great Lakes or serve to augment populations and reduce demographic stochasticity (Kelly et al. 2006). Some authors have argued that such genetic enhancements through admixture may increase the invasiveness of NIS (e.g. Kelly et al. 2006; Handley et al. 2011).

The mean abundance of zooplankton in Saint Lawrence River ballast water was significantly greater than that documented in a similar study of foreign exchanged ballast water bound for the Great Lakes (Bailey et al. 2011a; Briski et al. 2012b), indicating that domestic ballast water may currently be a more important vector for NIS in the Great Lakes than exchanged ballast discharged by transoceanic ships. Arguments that the Great Lakes—St. Lawrence River are a single contiguous ecosystem that should be managed as “common waters” are not supported by our analysis of biological similarity between source-recipient port-pairs. Finally, we provide evidence that species native to the St. Lawrence River, including some marine taxa, pose a risk for survival and future invasion of the Great Lakes.

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